

# Complementarity of two high-resolution spatiotemporal methods (hydroacoustics and acoustic telemetry) for assessing fish distribution in a reservoir

## *Complémentarité de deux méthodes à haute résolution spatio-temporelle (l'hydroacoustique et la télémétrie) pour évaluer la distribution spatiale des poissons dans un réservoir*

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**Abstract** – The complementarity of two high-resolution spatiotemporal acoustic methods, telemetry and hydroacoustics, was evaluated during the same time window to obtain fish distribution in a canyon-shaped reservoir, the Bariousses Reservoir (France). These methods act at an individual scale for telemetry and a community scale for hydroacoustics. The temporal scales are also different: telemetry offers continuous and long-term monitoring while a “snapshot” view is given by hydroacoustics. Day and night hydroacoustic surveys were carried out in this reservoir, during a 24-hour period in spring, using vertical and near-surface horizontal beaming. During this time window, 11 adult fish (length: 22–57 cm) from three species (roach, perch, and pikeperch) were tracked by telemetry. Four metrics were calculated with data collected by application of the two methods: distance to the nearest bank, distance to the tributary, fish depth, and bottom depth at the location. The contrasting (distance to the nearest bank, bottom depth) or partially similar results (distance to the tributary, fish depth) can be explained by the limitations associated with each method. The results obtained with telemetry are very sensitive to the species composition and the size of the tagged fish. The number of fish located in the epibenthic areas of the reservoir can be underestimated by hydroacoustics. This preliminary case study highlights that these methods act in a complementary way and their simultaneous use can provide better information on fish spatial distribution.

**Keywords** – fish; acoustic sampling methods; spatial distribution; lacustrine ecosystem.

**Résumé** – La complémentarité de deux méthodes à haute résolution spatio-temporelle : la télémétrie et l’hydroacoustique, a été examinée sur une même fenêtre temporelle, pour analyser la distribution des poissons dans le réservoir des Bariousses (France), un réservoir allongé en forme de canyon. Ces méthodes agissent à l’échelle de l’individu pour la télémétrie et à l’échelle de la communauté pour l’hydroacoustique. Les échelles temporelles sont aussi différentes : acquisition en continu sur de longues périodes en télémétrie et de manière instantanée pour l’hydroacoustique. Des parcours de jour et de nuit ont été réalisés dans cette retenue, sur une période de 24 heures, au printemps, en utilisant un sondeur vertical et horizontal. Durant cette période, 11 poissons de taille adulte (tailles allant de 22 à 57 cm) appartenant à trois espèces différentes (gardon, perche et sandre) ont été suivis par télémétrie. Quatre métriques ont été calculées à partir des données collectées par l’utilisation des deux méthodes : la distance à la rive la plus proche, la distance au tributaire, la position individuelle au sein de la colonne d’eau et la profondeur du fond correspondante. Les résultats contrastés (distance à la rive la plus proche, profondeur du fond) ou partiellement similaires (distance au tributaire, position individuelle au sein de la colonne d’eau) révélés par les analyses s’expliquent par les limites inhérentes aux méthodes. Les résultats obtenus par télémétrie sont très sensibles aux espèces et à la taille des poissons marqués. La fréquentation des zones épibenthiques peut être sous-estimée par l’hydroacoustique. Cette première étude de cas montre que ces méthodes agissent de manière complémentaire et qu’une meilleure description de la distribution spatiale des poissons peut être obtenue de par leur utilisation simultanée.

**Mots-clés** – poisson ; méthodes d’échantillonnage acoustiques ; distribution spatiale ; écosystème lacustre.

## 1 Introduction

Understanding the spatial distribution of fish populations remains a challenging fundamental issue in lakes ecology and a prerequisite to freshwater fishery management and conservation strategies (Cooke *et al.*, 2016).

The spatial distribution of fish within a waterbody is not random; fish exhibit spatial patterns related to the accomplishment of their vital functions – *i.e.*, reproduction, feeding, resting – requiring different environmental conditions (Lucas *et al.*, 2001).

Fish habitat selection depends on numerous abiotic and biotic factors that differ among species and ontogenic stages. First, fish search for a habitat with suitable conditions in terms of

temperature and oxygen concentrations (Fry, 1971; Kubečka & Wittingerova, 1998; Brosse *et al.*, 1999b; Lucas *et al.*, 2001; Järvalt *et al.*, 2005); second, food availability and predation drive their distribution (Savino & Stein, 1989; Gaudreau & Boisclair, 1998; Eklov & VanKooten, 2001; Gilliam & Fraser, 2001). Finally, during the reproductive period, physiological requirements are different and impact the distribution (Gillet, 2001).

To find optimum habitat conditions, fish complete circadian migrations in both directions: diel vertical migration (DVM) and diel horizontal migration (DHM) (Lucas *et al.*, 2001). DVMs are cyclic changes in the fish position in the water column, while DHMs are their movements between inshore and

offshore areas. However, the intensity and direction of migration depend on the ontogeny, species, and season. For instance, juveniles and adults generally migrate in opposite horizontal directions: juvenile fish perform night offshore migration (Romare *et al.*, 2003; Gliwicz *et al.*, 2006), whereas adults perform night inshore migration (Kubečka, 1993; Kubečka & Duncan, 1998; Zamora & Moreno-Amich, 2002; Jacobsen *et al.*, 2004; Říha *et al.*, 2011). Migration does not concern the entire population, and plasticity in the pattern of fish migration can occur (Eriksson, 1978; Busch & Mehner, 2012; Mehner & Kasprzak, 2011; Říha *et al.*, 2015).

The spatial distribution of the fish community is also structured along longitudinal and vertical gradients. In the majority of reservoirs, fish abundance and biomass decrease from the main tributary toward the dam (Brosse *et al.*, 1999a; Świerzowski *et al.*, 2000; Vašek *et al.*, 2003, 2004, 2006; Prchalová *et al.*, 2008) because of the riverine origin of fish fauna and a gradient of productivity (Vašek *et al.*, 2006). In mesotrophic or eutrophic reservoirs during spring and summer, fish are usually distributed in shallow depths, because of the attraction of warmer water and to avoid deoxygenated hypolimnion (Kubečka & Wittingerová, 1998; Čech & Kubečka, 2002; Vašek *et al.*, 2004). In cold and oligotrophic lakes, salmonid fish habitats are mainly in the deep cold water, below the thermocline (Guillard *et al.*, 2006; Yule *et al.*, 2013).

Improvements in technology have allowed scientists and managers of the waterbodies to perform more spatial behavioral fish studies (Lucas & Baras,

2000; Cooke *et al.*, 2016). All the methods dedicated to analyzing the spatial distribution of fish have intrinsic, environmental, and specific limitations. The spatiotemporal techniques available today are so diverse and with such high performances that most of the problems can be tackled by choosing the appropriate method or by combining tools and approaches (Lucas & Baras, 2000). Numerous studies have provided fish spatial distributions by coupling different methods (Tab. 1).

Telemetry, using electronic tags, is a capture-dependent method consisting in transmitting information to receivers (Cooke *et al.*, 2012), while hydroacoustics is a capture-independent method defined by the use of echosounding in water to measure the distribution and abundance of fish (Rudstam *et al.*, 2012); both methods provide high-resolution spatiotemporal data (Lucas & Baras, 2000; Arrhenius *et al.*, 2000; Belcher *et al.*, 2002; Rudstam *et al.*, 2012; Cooke *et al.*, 2012, 2013; Hussey *et al.*, 2015). Fish movements can be approached in four dimensions: horizontal (2D), vertical (depth, 3D), and over time (4D).

To our knowledge, probably because of the system perturbation or of the technical difficulty of implementation, there are only a few studies that used hydroacoustics and acoustic telemetry at the same time to describe fish spatial distribution. They can be considered as complementary methods allowing to gather better information on habitat use than each method separately. Telemetry helps locate individuals that were previously caught. Individuals are usually adults, and are limited in number. Hydroacoustics provides information

**Table 1.** Examples of studies using two or more different methods to study fish distributions in freshwaters.**Tableau 1.** Exemples d'études reportant l'utilisation de deux ou plusieurs méthodes différentes pour étudier la distribution des poissons dans les milieux d'eaux douces.

Methods	References
Hydroacoustics and netting	Baldwin & Polacek, 2011; Muška <i>et al.</i> , 2013
Hydroacoustics, gillnetting and others (scuba diving or ichthyoplankton net)	Imbrock <i>et al.</i> , 1996; Prchalová <i>et al.</i> , 2003,
Hydroacoustics and net towing	Kratochvil <i>et al.</i> , 2010
Hydroacoustics and radio telemetry	Grimardias <i>et al.</i> , 2017
Hydroacoustics and acoustic telemetry	Lyons & Lucas, 2002; McGrath <i>et al.</i> , 2003; Dunlop <i>et al.</i> , 2010
Acoustic telemetry and gillnetting	Smith <i>et al.</i> , 2011
Radio telemetry and mark-recapture	Auer, 1999
Radio or acoustic telemetry, PIT tags	Caswell <i>et al.</i> , 2004; Binder & McDonald, 2007
Mark-recapture and PIT tags	Caroffino <i>et al.</i> , 2009
Gillnetting and electrofishing	Mehner <i>et al.</i> , 2005
Beach seining, electrofishing and purse seining	Riha <i>et al.</i> , 2015

on the distribution of the whole fish community, without distinguishing between species (Lucas & Baras, 2000). The temporal scale of the study also differs: hydroacoustics is instantaneous, acting like a “snapshot”, whereas telemetry operates autonomously for extended periods (over a year), without additional maintenance need or battery change, allowing for continuous tracking (Klimley *et al.*, 1998; Heupel *et al.*, 2006; Baktoft *et al.*, 2012; McCauley *et al.*, 2014).

The main objective of this study was to test the complementarity (individual vs. community level) of these two methods to assess diel fish distribution during the same time window. The study was conducted in a canyon-shaped reservoir, the Bariousses Reservoir (France), with a main tributary. In this reservoir the depth variation and elongated morphology suggest an heterogeneous spatial distribution of organisms along the longitudinal axis, between littoral and pelagic zones, and a vertical distribution in the water

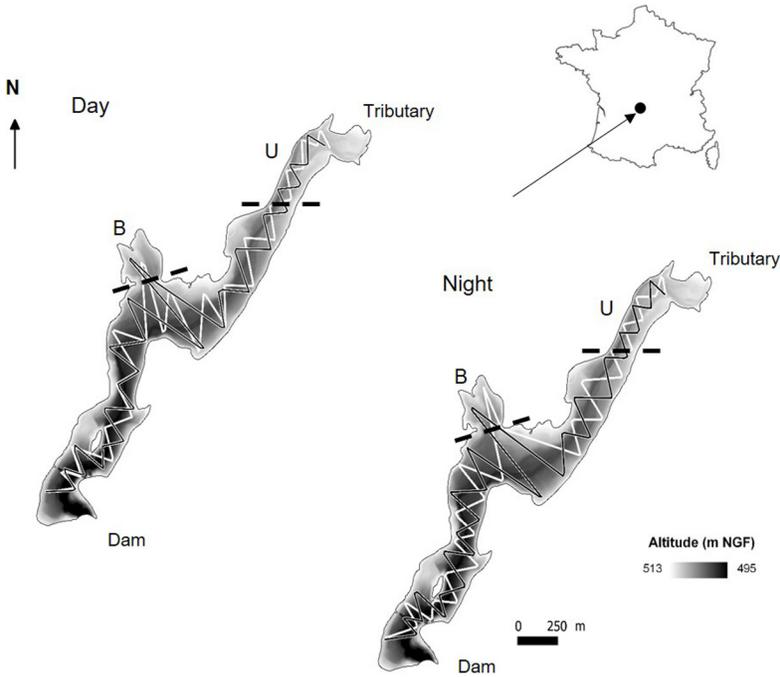
column (Duncan & Kubečka, 1995; Jurajda & Regenda, 2004).

## 2 Material and methods

### 2.1 Site description

The study was carried out in the Bariousses Reservoir, west central France (45.33°N, 1.49°E) (Fig. 1), an 80.9 ha impoundment of the Vézère River operated by Electricité de France (EDF). It is an elongated (3500 m long and 218 m wide at the mean water level), narrow lake (mean and maximum depths are 7.1 m and 18.9 m, respectively). The reservoir is monomictic, with a thermal stratification from summer to autumn (Roy, 2014).

Fish distribution in reservoirs is usually determined by the upstream-downstream gradient of the chemical and physical parameters (Brosse *et al.*, 1999a; Świerzowski *et al.*, 2000; Vašek *et al.*, 2003, 2004, 2006; Prchalová *et al.*, 2008). We defined two subareas with specific hydrology, substratum



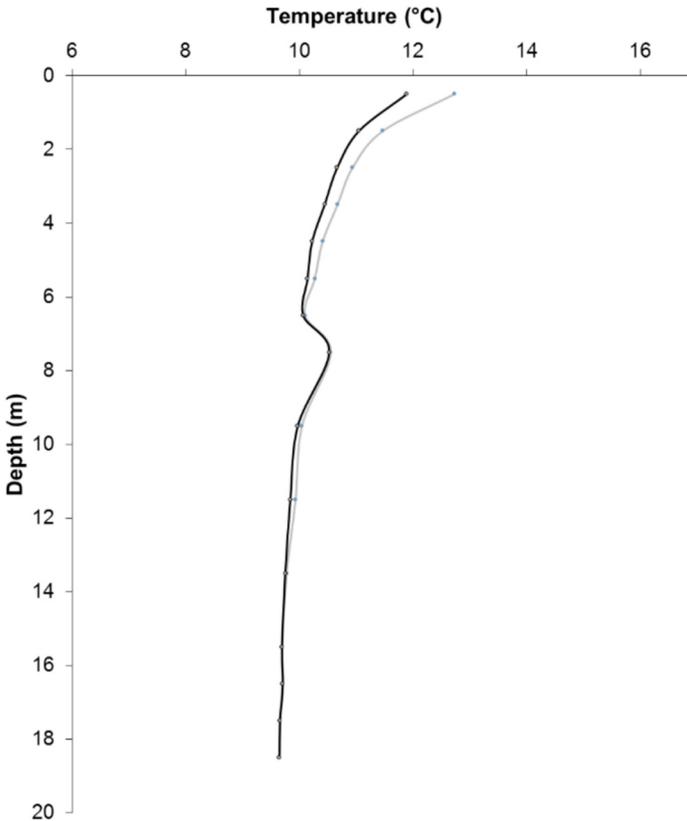
**Fig. 1.** Localisation and bathymetry (Altitude scale); day and night hydroacoustic zig-zag (white line: way-on; black line: way-back) and the two defined zones (U = upstream, B = Bay) and the tributary and dam positions. Maps were produced with QGIS 2.12.0, courtesy of EDF.

**Fig. 1.** Localisation et bathymétrie (échelle altitudinale); parcours en zig-zags effectués de jour et de nuit (ligne blanche : aller ; ligne noire : retour) et les deux zones délimitées (U = la partie amont, B = la baie) ainsi que les positions du tributaire et du barrage. Les cartes ont été réalisées avec QGIS 2.12.0 avec la permission d'EDF.

and low depth within the reservoir that may drive fish distribution: the upstream (sandy beach/mud with stumps) and the bay area (sandy beach/mud) (Roy, 2014) (Fig. 1).

During the survey, on 27 May 2013, the water level altitude was 512.5 m, corresponding to the 0.85 quantile over 2 years of measurements, which is high; a vertical profile of temperature was measured using a NKE thermometer installed in the deepest part of the reservoir (Fig. 2). A standardized gillnet survey (CEN, 2005) was performed at the end of August in 2010 (Roy, 2014),

which described the fish population of the reservoir. The reservoir is inhabited by 12 species, four of them represent the highest catch-per-unit effort: roach (*Rutilus rutilus*), ruffe (*Gymnocephalus cernuus*), perch (*Perca fluviatilis*), and pikeperch (*Sander lucioperca*) (Roy, 2014). The fish length (TL: total length) distribution obtained by gillnetting showed a main peak at 15 cm. Electro-fishing performed in the littoral zone, in spring 2011, revealed a fish length distribution mode of 10 cm (Fig. 3) (Roy, 2014). The reservoir has not been drained since 1997, and therefore



**Fig. 2.** Day (gray line) and night (black line) temperature profiles obtained on 27 May 2013. Each dot symbolizes the depth of an NKE thermometer.

**Fig. 2.** Profils de température obtenus de jour (ligne grise) et de nuit (ligne noire) le 27 mai 2013. Chaque point représente la profondeur d'un thermomètre NKE.

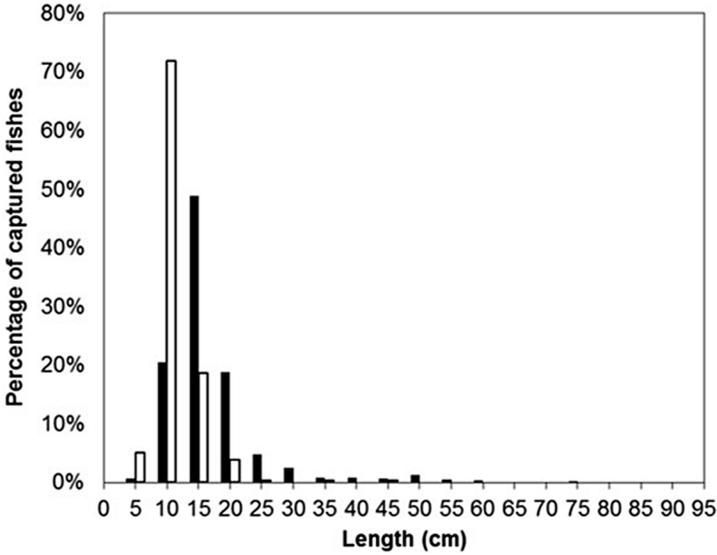
we assumed that the fish population has remained similar in terms of diversity and size composition.

## 2.2 Hydroacoustic survey

Hydroacoustic measurement was conducted on 27 May 2013 in the daytime and nighttime, at an approximate speed of  $8 \text{ km.h}^{-1}$ . Given the small size and morphology of the reservoir, day (15:19–17:16 coordinated universal time–CUT) and night (20:50–22:26 CUT) zig-zag trajectories

were used, on the way-on and way-back (two replicates) (Guillard & Vergès, 2007) (Fig.1). The trajectory was determined before the study to obtain representative data according to Aglen (1989). In our case, the degree of coverage, *i.e.*, the length of all transects divided by the square root of the reservoir area, is equal to 13.23 for the day survey and 13.12 for the night one.

A Simrad EK60 split-beam echosounder, 120 kHz frequency, controlled by the Simrad ER 60 (version 2.2.0)



**Fig. 3.** Total length distribution of fish obtained by gillnetting (black) (CEN, 2005) at the end of august 2010 and by electrofishing (white) in spring 2011.

**Fig. 3.** Distribution des longueurs totales des poissons capturés par pêche aux filets (en noir) (CEN, 2005) obtenue à la fin du mois d'août 2010 et par pêche électrique (en blanc) au printemps 2011.

program and connected to a GPS, was used for data acquisition. Two transducers, an elliptical one (ES 120-2.5  $\times$  10, nominal beam angle  $10^\circ \times 2.5^\circ$  at  $-3$  dB, beaming horizontally) and a circular one (ES 120-7C, nominal beam angle of  $7^\circ$  at  $-3$  dB, beaming vertically), were mounted on a platform on the side of the boat. The elliptical transducer was tilted  $3^\circ$  downward and the circular transducer beam was set 0.5 m below the surface. The pulse duration was 0.256 ms (Godlewska *et al.*, 2011), emitting four pulses per second, with power set at 100 W. Each transducer was calibrated once a year, using standard targets (Foote *et al.*, 1987). All acoustic data were analyzed from the echograms using post-processing software (Sonar5-Pro, version 6.0.3, Balk & Lindem, 2014). To exclude

the transducer nearfield and avoid the blind area close to the sounder (Simmonds & MacLennan, 2005), acoustic data within 2 m were excluded for the vertical beaming (Yule *et al.*, 2013) and data within 4 m were excluded for the horizontal beaming, as recommended by Draščík *et al.* (2009). As the reservoir was at the beginning of the seasonal stratification process (Fig. 2), the problem of beam bending is negligible (Simmonds & MacLennan, 2005). Data were processed up to the 50-m range for the horizontal beaming, corresponding to a water layer of approximately 4.8 m below the surface. With these settings, no target was simultaneously detected by both horizontal and vertical devices. For vertical beaming, a bottom 0.5 m layer was delimited to avoid the

inclusion of bottom detection in analyses and its accuracy was checked. All files were also checked for undesired non-fish echoes such as air bubbles, submerged macrophytes, debris, and buoys, and were deleted from the echograms (Emmrich *et al.*, 2012).

The hydroacoustic analysis was based on target counting, suitable for use when fish density is low, clear fish tracks are discernible within a noise, and when reliable GPS data are available (Drašík *et al.*, 2014). The latitude, longitude, and depth (m) of target positions were extracted from fish tracking.

For the vertical beaming, classic automatic tracking was applied. For the horizontal data, the cross-filter tracker (CFT) method (Balk & Lindem, 2014) was preferred instead of manual tracking, which is a labor-intensive and subjective process (Balk & Lindem, 2000), and instead of automatic tracking, which tends to generate fish-like tracks from noise echoes and to split tracks from fish. The CFT method improves automatic tracking in cases where single-echo detection (SED) echograms have low track quality combined with many noise-based detections. CFT encircles the echoes to be combined into track in an automatic way. The CFT used the cross-filter detector (CFD) (Balk, 2001) to improve track quality and to reduce erroneous detections in data with a low signal-to-noise ratio (SNR). For this survey, the CFD settings that provided the best results for tracking fish were for the detector (step1): Foreground filter=height 5 and width 1; Background filter=height 55 and width1; Offset +8 dB (Tušer *et al.*,

2009). For the evaluator (step 2), trace length and trace area were used. Minimum and maximum values for the trace length were 1–250 pings (Rakowitz *et al.*, 2008), and for the area, 8–400 samples in a detected region. The settings were chosen to find a compromise between rejected unwanted single echoes and to guarantee the maintenance of relevant parts of the fish traces containing a sufficient number of unaffected single echoes to size the fish properly. All fish tracks were manually checked.

For the vertical beaming, fish lengths were estimated using Love's (1971) equation (Emmrich *et al.*, 2012). To correct the target detection angle during mobile horizontal surveys, the deconvolution method is mainly used, based on a random distribution of fish (Kubečka *et al.*, 2009; Godlewska *et al.*, 2012), but does not allow access to the individual information, only to the length structure. However, this assumption is not true in the narrow parts of reservoirs, similar to a riverine environment, which then can lead to TS (target strength) over-estimations (Tušer *et al.*, 2009). In these parts, fish are mainly distributed with a side-aspect to the acoustic beam (Tušer *et al.*, 2009). The Bariousses Reservoir is smaller than the Rimov Reservoir (Czech Republic) where fish distribution was assessed to be in a non-random way and then positioned 90° to the beam. Owing to the riverine morphology of the Bariousses Reservoir, the fish distribution was considered to be generally positioned at 90° to the beam angle, and thus the horizontal side-aspect equation (TS side) for all the European fish species was used

(Eq. (1)) (Frouzová *et al.*, 2005). Each detected fish position was calculated using the following equation:

Equation (1):  $TS_{\text{side}} = a \log_{10} TL + b$ , where  $TS$  is in dB, and  $TL$  is the total length (mm),  $a$  is 24.71 and  $b$  is  $-89.63$ .

$TS$  was recorded by echosounding and  $TL$  was calculated from the inverse equation.

Thresholds were applied to both horizontal and vertical beaming for comparability of the two sampling modes. As the data were very noisy, the  $TS$  threshold was set at  $-50$  dB to avoid fish less than 5 cm in size; according to Love (1971), dorsal aspect regression (Simmonds & MacLennan, 2005) was chosen for vertical beaming, and according to Frouzová *et al.* (2005), side-aspect regression was chosen for horizontal beaming. This threshold was also chosen to avoid coarse suspended particles. Disproportionately strong echoes from acoustic phenomena were not identified as outliers by CFD (Rakowitz *et al.*, 2008). These outliers were removed so as not to oversize the fish, which were identified by a boxplot graphical technique (Tukey, 1977).

Prior to the hydroacoustic survey, we checked that the hydroacoustic system did not affect the telemetry system or make it collapse because of the possible interferences between the echosounder pulse emission and the tags of the telemetry system.

## 2.3 Telemetry survey

The telemetry dataset was considered on the same time windows as the hydroacoustics.

### 2.3.1 Tracking system

An array of 40 underwater VR2W 69-kHz omnidirectional acoustic receivers (VEMCO) was deployed in January 2012 throughout the reservoir. The receiver deployment took into account the bathymetry, the shape of the reservoir, and the maximum intensity of water-level fluctuations to avoid the receivers being beached. The receivers were positioned at an average of 150 m apart (range was 72–223 m) and at an average depth of 6 m (range from 2 m to 15 m) (Roy *et al.*, 2014).

The VEMCO Positioning System (VPS) (Smith, 2013) was used to calculate the fish positions in the horizontal plane. Only positions with an HPE (Horizontal Position Error, a parameter provided by the VPS, Smith, 2013) less than 20 were retained to filter false locations. This corresponded to about 86% of the full position dataset with a mean error of 3.5 m (Roy *et al.*, 2014). The depth of the VPS positions was measured with a pre-calibrated hydrostatic pressure sensor (accuracy: 2.5m).

### 2.3.2 Fish tagging

A total of 143 fish, mainly adults (22.0–62.9 cm), were tagged during January 2012–April 2013 period with VEMCO V9P-2L or V8-4L acoustic transmitters (mean interval burst of 90 or 120 s) in the context of a fish habitat analyze (Roy, 2014). Fish were captured in Bariousses Reservoir with gillnets set at dawn, day, and dusk for maximally 2 hours or by specialist anglers. Due to the low catch of pike-perch, the sample was completed with fish from an extensive pond aquaculture

and introduced into the reservoir after marking. Once caught, fish were individually anesthetized, and the transmitter was surgically inserted in the peritoneal cavity. Then, fish were placed in an oxygenated tank to recover during at least a few hours to one night which also enabled to highlight abnormal behaviours (details in [Roy, 2014](#)) before being released close to their capture site. Not to bias the analysis with behaviors that could be linked to the surgery, only positions recorded at least 2 days after the release were retained ([Bridger & Booth, 2003](#)).

Given the theoretical battery lifetime, fish mortality and the definitive exit of some individuals from the detection zone, only 21 of the 143 fish were potentially detectable during the hydroacoustic campaign. Individual tracked during the time window of the hydroacoustic survey and belonging to the most represented species in the fish community according to gillnetting ([Roy, 2014](#)) were retained in the analysis *i.e.* three roach, six perch, and two pikeperch. Due to the low number of individuals detected during the hydroacoustic campaign, the species analysis was performed for illustrative purposes only.

## 2.4 Spatial data

The fish spatial distributions were described using R 3.3.1 statistical software ([R Development Core Team, 2016](#)). All the fish positions retained in the analysis were mapped. The percentage of fish positions in two areas, the upstream part and the bay part, was calculated for the two methods. To obtain the bottom depth at the fish positions, the reservoir was discretized

into squares of  $10\text{ m} \times 10\text{ m}$  and the percentage of the positions per square was calculated. For these squares, the bottom depth was calculated with bathymetric data measured *via* a multibeam sounder giving a  $2\text{ m} \times 2\text{ m}$  resolution map (EDF). When the position of the square center was discriminated to be out of the water surface, the depth was approximated by replacing the missing value with the mean bottom depth of the 5% lowest values of the water column height. The Euclidian distance of individual fish to the bank and to the tributary was calculated with the *rgeos* package ([Renard & Bez, 2005](#)).

## 2.5 Split of fish community by length

The fish marked by telemetry were mainly adults, whereas those recorded by hydroacoustics covered all the range size of the community. As fish have different ecological preferences, but also behavior, based not only on species but also on ontogeny ([Riha \*et al.\*, 2015](#)), especially between juveniles and adults, the fish detected by hydroacoustics were split into two groups: fish  $< 20\text{ cm}$  and fish  $\geq 20\text{ cm}$ .

## 2.6 Fish positioning and statistical analyses

Four metrics related to fish position were calculated using the two datasets (hydroacoustics and telemetry): distance between the individual position and the nearest bank, distance to the tributary, fish depth, and bottom depth at the location. The fish distribution from hydroacoustic data was analysed for small and large fish groups, and for

**Table 2.** Number of hydroacoustic-tracked fish obtained with horizontal and vertical beaming during the day and night survey. The number of fish with TL < 20 cm and TL ≥ 20 cm is also shown.

**Tableau 2.** Nombre de poissons détectés en hydroacoustique avec les sondeurs horizontaux et verticaux pendant les campagnes de jour et de nuit. Le nombre de poissons avec une LT < 20 cm et LT ≥ 20 cm est aussi présenté.

	Horizontal		Vertical		Total	
	Day	Night	Day	Night	Day	Night
All fish	489	563	23	16	512	579
Fish < 20 cm	357	461	22	13	379	474
Fish ≥ 20 cm	132	102	1	3	133	105

the telemetry data the fish distribution was set according to species identification. All distributions being not normal, Wilcoxon non-parametric tests were used to compare the different spatial metrics.

The statistical analyses were performed with R 3.3.1 statistical software (R Development Core Team, 2016).

### 3 Results

A total of 1,091 individual fish (Tab. 2) were detected during the day and the night surveys *via* the hydroacoustic method, mainly recorded by horizontal beaming (96% of detections). The method showed a 13% increase in the number of fish detected during the night owing to an increase in the detection of small fish compared with large fish that are detected less often.

By using telemetry, 361 individual positions from the three roach, six perch, and two pikeperch were recorded during the hydroacoustic survey period (Tab. 3). At night, the total number of recorded positions decreased (two perch not detected).

Fish length distributions obtained during the day and night from hydroacoustic surveys (Fig. 4) showed a

mode at 10 cm. They were analogous to the one obtained by electrofishing (Fig. 3) performed during a similar period. The mode obtained by gillnetting survey performed later, in autumn, was higher by only 5 cm.

#### 3.1 Global fish distribution

Fish distributions (number of individual positions recorded by hydroacoustics and by telemetry) were shown using repartition maps (Fig. 5). With hydroacoustics, fish were detected in all parts of the lake, *i.e.*, in littoral and pelagic zones of the reservoir. During the day, a higher percentage of fish was observed close to the tributary in the upstream part of the lake (35.9% during the day and 10.9% during the night). At night, the highest proportion of fish (22.4%) was recorded in the bay, in the intermediate part of the reservoir.

Telemetry positions were less homogeneously distributed than with hydroacoustics. Indeed, no position was recorded by the telemetry method in the center of the reservoir. During the day, the telemetry positions were scattered along and close to the left bank and no fish was observed in the upstream part, while during the night

**Table 3.** Total length (TL, cm) and number of individual positions of fish recorded by the telemetry method during the hydroacoustic surveys. The presence (yes) or absence (no) of a pressure sensor allowing the fish position in the water column to be defined is noted.

**Tableau 3.** Longueur totale (LT, cm) et nombre de positions individuelles de poissons enregistré par télémétrie pendant la campagne d'hydroacoustique. La présence (oui) ou l'absence (non) d'un capteur de pression permettant de déterminer la position du poisson au sein de la colonne d'eau est renseignée.

Transmitter (mean burst interval (s), type of transmitter)	Species	Pressure sensor	Total length (cm)	No. Posit. Day	No. Posit. Night
T123 (120, V8-4L)	roach	no	25.4	27	26
T124 (120, V84L)	roach	no	23.6	7	12
T125 (120, V8-4L)	roach	no	22	17	14
T56 (90, V9P-2L)	perch	yes	43	44	/
T112 (120, V9P-2L)	perch	yes	48.6	18	18
T114 (120, V9P-2L)	perch	yes	37.9	1	/
T115 (120, V9P-2L)	perch	yes	41.5	26	15
T117 (120, V9P-2L)	perch	yes	32.2	20	10
T118 (120, V9P-2L)	perch	yes	40.5	15	38
T15 (90, V9P-2L)	pikeperch	yes	42.8	4	1
T111 (120, V9P-2L)	pikeperch	yes	57	25	23
			Total	204	157

one individual was detected in this part and accounted for 24.2% of the telemetry positions. Two fish were positioned in the bay at daytime and one during the night (accounted for 31.86% of positions during the day and 14.68% during the night).

### 3.2 Distance to the tributary

The results obtained with the two methods were different according to the period considered.

During daytime, distances to the tributary estimated by telemetry were higher than those calculated with hydroacoustics, whatever the size group considered ( $p < 0.05$ ) (Fig. 6).

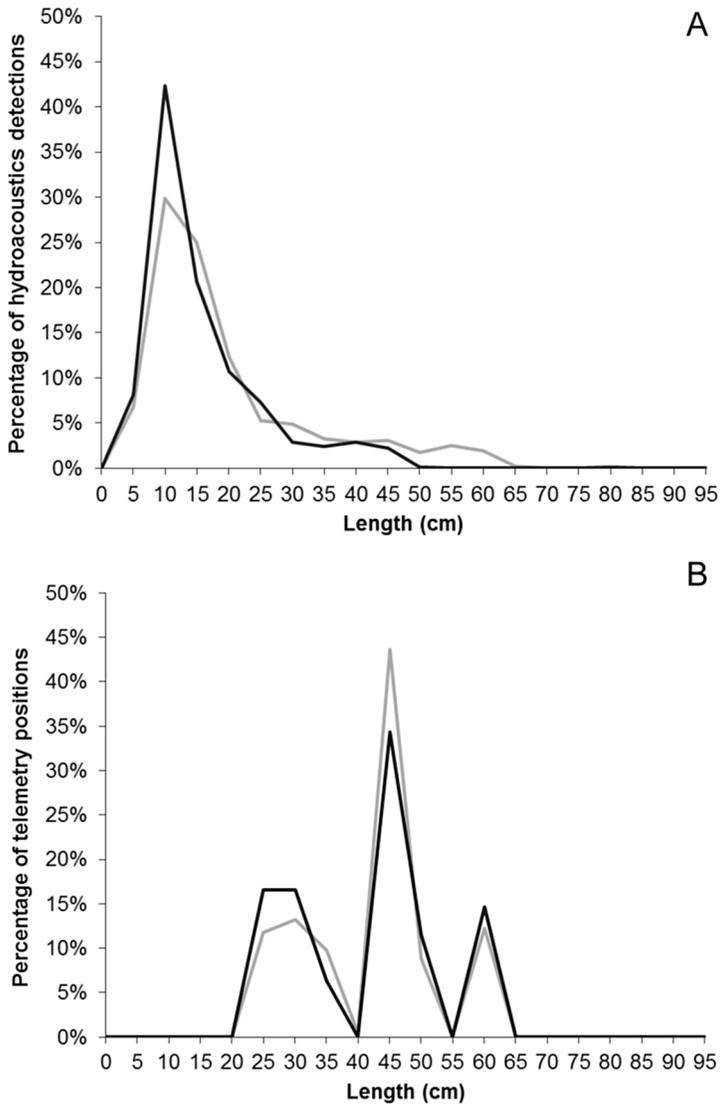
During nighttime, distances calculated with the whole hydroacoustic dataset and for hydroacoustic data of fish smaller than 20 cm were significantly higher than those estimated with telemetry. Conversely, the distribution of fish larger than 20 cm assessed with hydroacoustic survey did not differ to

the distribution of fish obtained with telemetry ( $p > 0.05$ ).

The diel direction pattern also differed; the telemetry positions were further away from the tributary in the day than at night (mean<sub>day</sub>: 1,568.0 m  $\pm$  24.7; mean<sub>night</sub>: 1,300.4 m  $\pm$  51.3). The opposite was shown with hydroacoustics; statistical tests confirmed a higher fish density in the areas close to the tributary in the daytime (Fig. 6) ( $p < 0.05$ ; mean<sub>day</sub>: 1,090.4 m  $\pm$  31.8; mean<sub>night</sub>: 1,449.1 m  $\pm$  24.7) than during the night.

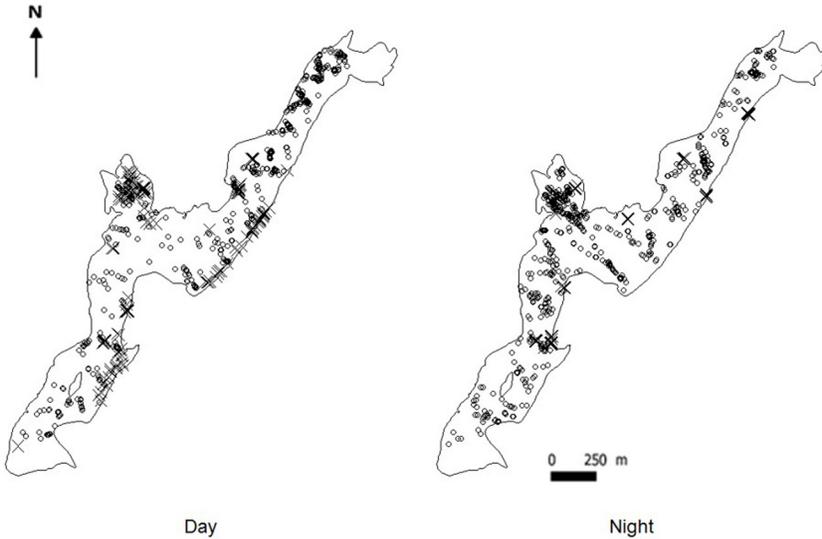
There was no significant difference in the distribution of hydroacoustic-detected fish sized  $< 20$  cm and  $\geq 20$  cm, in the day nor at night, even if the distribution of fish  $\geq 20$  cm was in both cases (day and night) more spread out close to the tributary.

Regarding the distance to the tributary of the different species, only perch showed a significant diel pattern by being closer to the tributary during the day than at night (Tab. 4,  $p < 0.05$ ).



**Fig. 4.** Length distribution of (A) fish detected by hydroacoustics and (B) by telemetry during day (grey) and night (black) on 27 May 2013.

**Fig. 4.** Distribution en taille (A) des poissons obtenue à partir des données d'hydroacoustique et (B) obtenue à partir des positions en télémétrie pendant le jour (gris) et la nuit (noir) le 27 mai 2013.



**Fig. 5.** Day and night spatial distribution of fish in the Bariousses Reservoir in May 2013 obtained with hydroacoustics (empty black circles) and telemetry (black crosses) methods.

**Fig. 5.** Distribution spatiale des poissons dans le réservoir des Bariousses obtenue de jour et de nuit par hydroacoustique (cercles noirs vides) et télémétrie (croix noires).

### 3.3 Distance to the bank

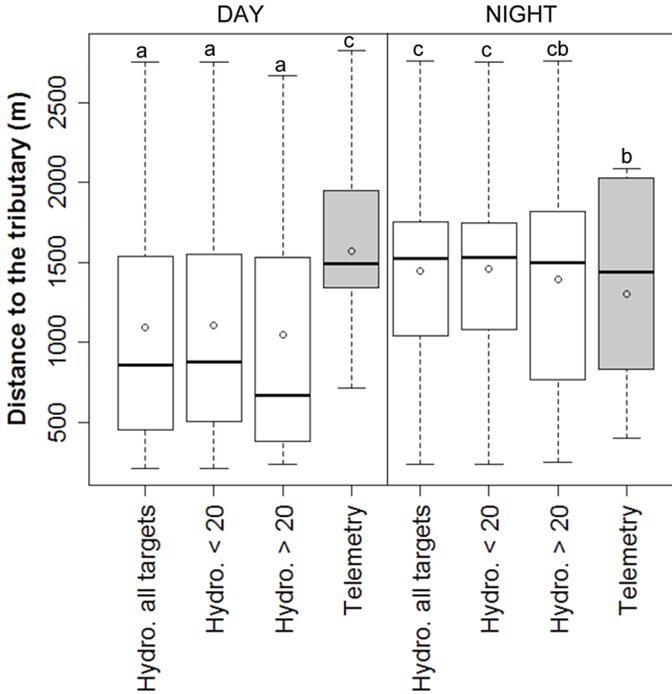
Statistical analysis showed that fish positions obtained with the telemetry method were closer to the bank than the ones obtained with hydroacoustics, for all the datasets considered ( $p < 0.05$ ) (Fig. 7). The two methods provided opposite diel patterns. Indeed, with telemetry, a greater distance was observed between the fish and the bank during the day compared with the night (mean<sub>day</sub>:  $45.5 \text{ m} \pm 2.1$ ; mean<sub>night</sub>:  $38.9 \pm 2.8$ ), and the opposite was observed for hydroacoustics except in the case of large fish. The distribution of fish larger than 20 cm detected with this method did not differ significantly between day and night (mean<sub>day</sub>:  $62.0 \text{ m} \pm 2.33$ ; mean<sub>night</sub>:  $65.6 \text{ m} \pm 3.1$ ), even if the distribution was more spread out toward the bank at night.

Global species diel pattern for marked individual was different: Perch were closer to the bank at nighttime while roach moved in the opposite direction (Tab. 3,  $p < 0.05$ ).

### 3.4 Depth

Most of fish positions detected with the two methods were located in the upper water layer ( $< 6 \text{ m}$  depth) and similar patterns for the diel distributions were found with the two methods; fish were deeper during the day than at night ( $p < 0.05$ ) (Fig. 8).

Differences in depth distribution were highlighted. Most of fish detected with the hydroacoustic method were closer to the surface (for all fish: mean<sub>day</sub>:  $2.2 \text{ m} \pm 0.1$ , mean<sub>night</sub>:  $1.5 \text{ m} \pm 0.1$ ) (Fig. 8) compared with the



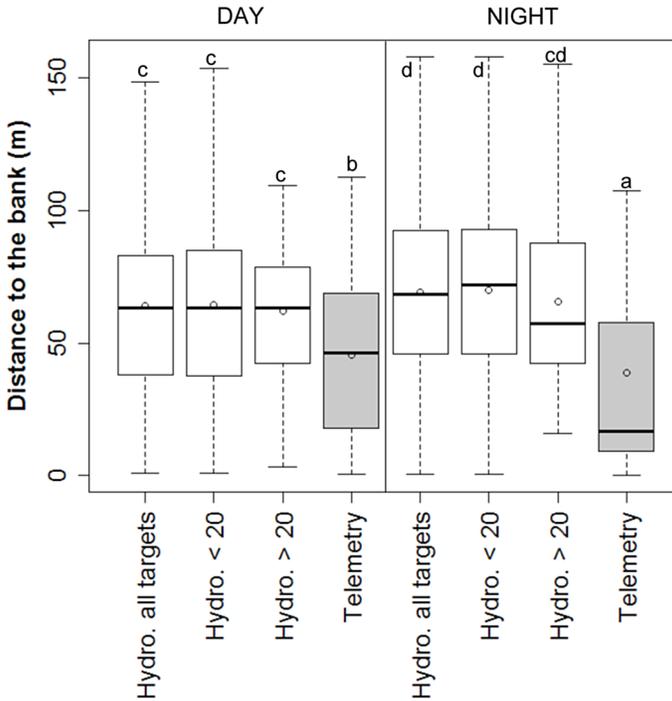
**Fig. 6.** Boxplots of the fish distance to the tributary obtained with hydroacoustics (“hydro.”) and telemetry surveys during the day and at night. Median = dark horizontal line in bold; boxes represent 25th and 75th percentiles. Horizontal lines = maximum and minimum values. Black circles = mean values. The results of Wilcoxon test are given (variables that do not share the same letter are significantly different).

**Fig. 6.** Boîtes à moustaches de la distance des poissons au tributaire obtenues en hydroacoustique (« hydro. ») et en télémétrie pendant le jour et la nuit. Médiane = ligne noire horizontale en gras ; les boîtes représentent les 25 et 75<sup>ème</sup> percentiles. Lignes horizontales = valeurs maximale et minimale. Cercles noirs = valeurs moyennes. Les résultats du test de Wilcoxon sont donnés (les variables qui ne partagent pas la même lettre sont significativement différentes).

**Table 4.** Mean  $\pm$  standard error distance (m) to the tributary, to the bank, and depth for each detected species by telemetry. No data are available for roach depth. The number of individuals is in parentheses.

**Tableau 4.** Moyenne  $\pm$  erreur standard de la distance (m) au tributaire, à la rive et profondeur de chaque espèce détectée par la télémétrie. Les données concernant la profondeur ne sont pas disponibles pour le gardon. Le nombre d’individus est indiqué entre parenthèses.

		Roach	Perch	Pikeperch
Distance to the tributary (m)	Day	1,552.0 $\pm$ 81.0 (3)	1,593.9 $\pm$ 35.8 (6)	1,485.2 $\pm$ 19.6 (2)
	Night	1,578.1 $\pm$ 74.5 (3)	1,078.1 $\pm$ 79.8 (4)	1,449.1 $\pm$ 6.5 (2)
Distance to the bank (m)	Day	69.5 $\pm$ 3.9 (3)	34.6 $\pm$ 2.6 (6)	50.1 $\pm$ 0.5 (2)
	Night	81.6 $\pm$ 2.5 (3)	8.94 $\pm$ 0.8 (4)	47.4 $\pm$ 1.6 (2)
Depth (m)	Day	/	4 0. $\pm$ 1 (6)	6.4 $\pm$ 0.2 (2)
	Night	/	2.5 $\pm$ 0.1 (4)	5.8 $\pm$ 0.2 (2)



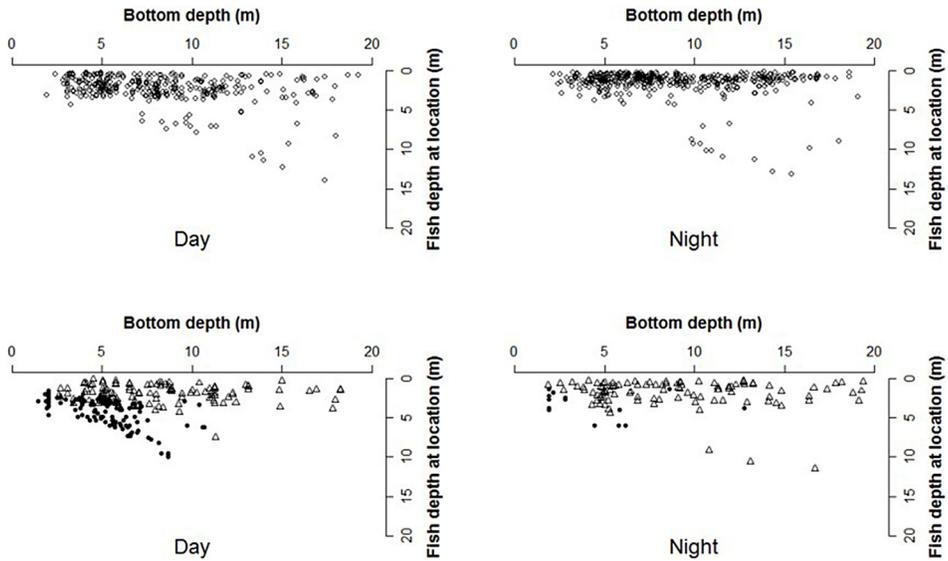
**Fig. 7.** Boxplots of the fish distance to the bank from hydroacoustics (“hydro.”) and telemetry surveys during the daytime and at night. Median=dark horizontal line; boxes represent 25th and 75th percentiles. Vertical lines=maximum and minimum values. Red circle=mean values. The results of Wilcoxon test are given (variables that do not share the same letter are significantly different).

**Fig. 7.** Boîtes à moustaches de la distance des poissons à la rive obtenues en hydroacoustique (« hydro. ») et en télémétrie pendant le jour et la nuit. Médiane = ligne noire horizontale en gras ; les boîtes représentent les 25 et 75<sup>ème</sup> percentiles. Lignes horizontales = valeurs maximum et minimum. Cercles noirs = valeurs moyennes. Les résultats du test de Wilcoxon sont donnés (les variables qui ne partagent pas une même lettre sont significativement différentes).

depth distribution obtained *via* telemetry (mean<sub>day</sub>:  $\pm 0.1$ , mean<sub>night</sub>:  $3.2 \pm 0.2$ ). In addition, fish followed with telemetry were located in epibenthic habitats whereas the hydroacoustic-detected targets were close to littoral/surface areas. The intensity of the diel shift was more pronounced for the hydroacoustic targets with no distinction between length and for fish smaller than 20 compared with the distribution of fish larger than 20 from telemetry and

acoustics data. The < 20 cm and  $\geq 20$  cm distribution of hydroacoustic-detected fish did not significantly differ in the day but there were differences at night.

Compared to perch distribution in the water column, pikeperch stayed deeper whatever the time of day considered (Tab. 4,  $p < 0.05$ ). Only perch showed a significantly different pattern between day and night ( $p < 0.05$ ).



**Fig. 8.** Day (left) and night (right) depth plots indicating the fish position depth and bottom depth obtained with telemetry (black circles) and hydroacoustics (empty items). For the hydroacoustic method, the upper plots represent the positions of fish with a length of  $< 20$  cm (white circles) and the bottom plots represent the positions of fish with a length of  $\geq 20$  cm (white triangles).

**Fig. 8.** Graphique représentant la profondeur des poissons et la profondeur du fond à la position des poissons, pour le jour (à gauche) et la nuit (à droite) en télémétrie (cercles noirs pleins) et en hydroacoustique (cercles blancs). Pour la méthode hydroacoustique, les graphiques du haut représentent les positions pour les poissons avec une taille  $< 20$  cm (cercles blancs) et les graphiques du bas représentent les positions pour les poissons avec une taille  $\geq 20$  cm (triangles blancs).

## 4 Discussion

### 4.1 Fish horizontal distributions

In elongated reservoirs, during summer, a longitudinal gradient is generally observed with a higher fish abundance in the upstream part (main tributary area) (Pont & Amrani, 1990; Urabe, 1990; Fernando & Holčík, 1991; Brosse *et al.*, 1999a; Świerzowski *et al.*, 2000; Vašek *et al.*, 2003, 2004). This was also observed the day with hydroacoustics in our study site. Different hypotheses about this common gradient were posited by Vašek *et al.* (2004). The first one states that because fish fauna in reservoirs have a riverine origin, they are not completely adapted to lacus-

trine conditions and find their habitats in shallow inshore areas, close to the tributary in the upstream part (Fernando & Holčík, 1991). Second, the upstream part is generally more productive (Straskraba, 1998) with a gradient of chlorophyll-a concentrations and zooplankton densities from the tributary to the dam (Urabe, 1989; Fernandes-Rosado *et al.*, 1994; Dohet & Hoffmann, 1995; Fernandes-Rosado & Lucena, 2001; Vašek *et al.*, 2003). As a result, zooplanktivorous fish have a higher density in this area and their distribution may reflect the longitudinal gradient of productivity: Urabe (1990) and Siler *et al.* (1986) reported that the abundance of planktonivorous fish

during summer decreased from the tributary to the dam. This is probably what we observed in our study. Indeed, the community is dominated by roach (Roy, 2014) and it was shown that, in conditions similar to those observed in the Bariousses, this species is able to forage most exclusively on crustacean zooplankton (Vašek *et al.*, 2003).

Hydroacoustics data revealed that the distances of the fish to the tributary and to the bank are greater at night than during the day. Small fish are generally associated with a structured habitat within the littoral during daytime (Lewin *et al.*, 2004; Gliwicz *et al.*, 2006), and at dusk small fish migrate to the pelagic zone where zooplankton prey are more abundant (Romare *et al.*, 2003; Gliwicz *et al.*, 2006).

As the tributary area is a shallower part, we can hypothesize that at dusk zooplanktivorous fish migrate to open waters to follow food supply (Bohl, 1980; Romare *et al.*, 2003; Gliwicz *et al.*, 2006) and to reduce the risk of predation as light intensity decreases (Cerri, 1983).

The telemetry method did not reveal a significant difference in the diel distribution to the tributary, whereas the opposite distribution was seen during the day with a greater distance to the tributary and to the bank. Only three roach were recorded by telemetry during the hydroacoustic survey; however, one individual showed the classic diel pattern of zooplanktivorous fish. The majority of tagged fish were perch and the distribution obtained by telemetry may reflect the spatial distribution of this species. Perch switch predominantly to piscivory when they reach two years of age and they exploit the open

water zone (Parker *et al.*, 2009). Unlike cyprinids, perch are efficient competitors and predators in clear water (Diehl, 1988; Radke & Gaupisch, 2005) and this species has a higher biomass in less productive, downstream areas (Vašek *et al.*, 2016). Perch swim continuously parallel to the bank during the day and get close to the littoral zone to rest at night, exhibiting routine homing behavior (Zamora & Moreno-Amich, 2002).

The distribution of the large-sized group was closer to the one obtained with telemetry at night, and similar results were found by Lyons and Lucas (2002) in The River Trent. At night, the majority of fish are dispersed in the water column, whereas during daytime fish aggregate in schools or are close to the bottom making them less accessible to acoustics methods. Consequently, echosounding cannot easily be used to quantify the fish distribution during daytime (Duncan & Kubečka, 1993; Kubečka & Wittingerova, 1998; Ye *et al.*, 2013).

## 4.2 Fish depth

The two methods showed convergent patterns with fish mainly located in the warmer water (< 6 m deep). Vertical beaming alone underestimated the total amount of fish in Bariousses Reservoir by 96%. This result is in agreement with other hydroacoustic studies where the exclusive use of vertical beaming has led to underestimate fish density by 5100% (Kubečka & Wittingerova, 1998; Knudsen & Sægrov, 2002; Djemali *et al.*, 2009). In thermally stratified reservoirs, fish densities or biomass sampled with horizontal beaming are higher than those determined with vertical beaming

(Kubečka & Wittingerova, 1998; Draštík *et al.*, 2009). In the surface layer, the fish population is virtually undetectable when using only vertical beaming owing to the near field of the transducer. During our survey the process of thermal stratification had just begun; however, the slight difference in temperature observed between layers was sufficient to drive the spatial distribution of the fish. These results confirm the importance of horizontal beaming for assessing the spatial distribution of fish in thermally stratified reservoirs. A typical diel vertical shift with an ascent at dusk and a descent at dawn was also revealed with the two methods.

This distribution is explained by the multifactorial hypothesis of the “anti-predation window” (Clark & Levy, 1988) and also by the thermal niche hypothesis for zooplanktivorous fish. During the day, prey fish find refuge in deeper zones with darker conditions and move within this antipredation window. To minimize the cost of swimming (Ohlberger *et al.*, 2008) and speed up the digesting rate during the non-feeding phase at night (Wurtsbaugh & Neverman, 1988; Neverman & Wurtsbaugh, 1994), fish find temperatures close to their preferendum (Mehner *et al.*, 2010). For predators, such as perch that dominated the telemetry dataset, the risk of predation is not an issue (Mehner, 2012). The foraging and bioenergetics hypotheses have also been most successful at explaining DVM (Bevelhimer & Adams, 1993).

Predator avoidance and feeding opportunities should explain the distribution of small planktivorous fish that stay in deeper layers during the day. At night, differences are highlighted when fish

distribution is driven by bioenergetic efficiency and when each ontogenic stage seeks optimum temperature layers.

In our study, fish from the telemetry dataset were found to be deeper than fish detected by hydroacoustics. Apart from measurement uncertainty (2.5 m), fish seeking energetically optimum temperatures could explain the difference (Mehner *et al.*, 2010). Perch dominated the telemetry sample but the community is dominated by roach. The location of perch in deeper layers compared with roach has been reported in numerous studies in lakes and reservoirs (Persson, 1986; Horppila *et al.*, 2000; Kahl & Radke, 2006).

Stronger diel differences in fish depth were observed for small size fish in hydroacoustics compared with large size fish using the two methods implemented here. Ontogenic differences in the thermal niche of fish (Portner & Farrell, 2008) could explain the difference in the observed amplitude of migration.

#### 4.3 Methodological considerations

This study highlights the differences between results provided by the two methods that can be interpreted by the different biological scale (community or individual), as discussed in the previous section and temporal scale (punctual or continuous). Hydroacoustics provides an image of the repartition of the fish community during the survey, over a short time scale, whereas telemetry reflects the detailed tracks of some individuals, adults in our case, detected during the survey.

We have shown that the importance of the upper part of the reservoir highlighted by the hydroacoustic survey can be

underestimated with telemetry data, even if specimens of the dominant fish species are included in the survey. The results of telemetry were highly dependent on the species and the number of tagged fish that were considered. Major disadvantages of this method are the cost of the system and the burdensome tagging procedure that often limits the number of fish tracked. Atypical movement can greatly affect telemetry results when a small number of fish are tagged. Some individuals can move a great distance, for example, when seeking for a new home range (Ebner & Thiem, 2009), and a proportion of moving individuals have been reported for perch (Zamora & Moreno-Amich, 2002). Therefore, for the various metrics discussed in the previous part, results obtained in telemetry are based on a small number of individuals and are presented here for information only. Results of the two methods would probably be in better agreement if the sample of tagged fish was more important, more representative of the composition of the whole community and the ecospecies dominance in the system. In the future, the use of micro-transmitters will make this easier.

Environment plays also a major role in the efficiency of the system (Gjelland & Hedger, 2013; Kessel *et al.*, 2014, 2015; Ottera & Skilbrei, 2016) and needs to be estimated, which was done in the present study (Roy *et al.*, 2014) but is generally still uncommon.

Conversely, regarding the depth distribution obtained with telemetry, the use of the epibenthic habitat is probably undervalued by hydroacoustics. Indeed, fish close to the bottom cannot be easily discriminated from bottom echoes and submerged mac-

rophyte or tree roots also cause difficulties in the use of the hydroacoustic technique in shallow waters. However, the total volume sampled by this method is still very large and high-resolution spatial records of fish distribution can be created to inform on the fish distribution at the community level.

In addition, the method does not allow for the determination of species composition and must be complemented by other techniques: trawling, purse seining, and gillnetting are commonly used (Parkinson *et al.*, 1994; Yule, 2000; Baldwin & McLellan, 2008; Winfield *et al.*, 2009; Yule *et al.*, 2013). Studying the distribution of different size-classes has other limitations. Fish sizing is relatively simple with vertical echosounding because the fish are viewed from above and appropriated relationships are generally available (Love, 1977; Foote *et al.*, 1987; Simmonds & MacLennan, 2005). With horizontal mobile beaming, the angle of the fish position to the beam axis is unknown, and then the conversion to length is difficult (Godlewská *et al.*, 2012). Deconvolution does not provide information on the individual position and subsequent attributed size. We hypothesize that fish are not randomly distributed but oriented at 90° to the acoustic axis, because of the small width of the Bariousses Reservoir. However, deviation biased the distribution, and the number of large individuals is probably underestimated.

Conversely, in noisy environments, small fish can also be underestimated (Draštík *et al.*, 2009). However, in this study, the length distribution obtained by gillnetting and electrofishing is close

to the one obtained by hydroacoustics. By selecting the fish size, the distribution pattern becomes similar. These results are encouraging for future studies on fish distribution taking into account ontogeny.

Even if the estimation of fish size is still a challenging limitation in the hydroacoustics horizontal scan, this study confirms that, in this type of reservoir (shallow, elongated and monomictic) at the start of the thermal stratification, horizontal beaming is crucial to study fish distribution (Kubečka & Wittingerova, 1998; Knudsen & Sægrov, 2002; Draštk et al., 2009).

To conclude, the spatial distribution of fish in an elongated reservoir has the potential to be better described, using two high spatiotemporal methods – telemetry and hydroacoustics – in parallel to complement each other. Hydroacoustics gives a “snapshot” at the community level and telemetry gives continuous data at the individual and species level. The differences in the results obtained could be limited by tagging a more representative sample of the community in terms of sizes, with the use of microtransmitters, and species and by improving the detection of fish in epibenthic areas with hydroacoustic data acquisition. More experiments are needed with several time and space repeated echosounding to improve robustness (different sampling in a similar environment within a season) and to allow better generalization of the results (different sampling in other sites, at different seasons). However, we can draw preliminary conclusion about the utility of the complementation of these two high spatiotemporal acoustic methods for assessing fish spatial distribution in a reservoir and proposed

typical metrics to do that. In using these two methods simultaneously, new knowledge is provided that could be very useful for fish management (Prado & Pompeu, 2014).

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## REFERENCES

- Aglen A., 1989. Empirical results on precision-effort relationships for acoustic surveys. *ICES CM B* 30: 28 p.
- Arrhenius F., Benneheij B.J., Rudstam, L. G. & Boisclair D., 2000. Can stationary bottom split-beam hydroacoustics be used to measure fish swimming speed *in situ*? *Fish. Res.* 45(1): 31–41.
- Auer N.A., 1999. Population characteristics and movements of lake sturgeon in the Sturgeon River and Lake Superior. *J. Great Lakes Res.* 25(2): 282–293.
- Baktoft H., Aarestrup K., Berg S., Boel M., Jacobsen L., Jepsen N., ... & Skov C., 2012. Seasonal and diel effects on the activity of northern pike studied by

- high-resolution positional telemetry. *Ecol. Freshw. Fish* 21(3): 386–394.
- Baldwin C.M. & McLellan J.G., 2008. Use of gill nets for target verification of a hydroacoustic fisheries survey and comparison with kokanee spawner escapement estimates from a tributary trap. *North Am. J. Fish. Manage.* 28(6): 1744–1757.
- Baldwin C.M. & Polacek M., 2011. Abundance and seasonal shifts in vertical and horizontal distribution of lake whitefish (*Coregonus clupeaformis*) in a western United States reservoir. *J. Freshwater Ecol.* 26(2): 171–183.
- Balk H., 2001. Development of hydroacoustic methods for fish detection in shallow water. PhD thesis, Faculty of Mathematics and Natural Science, University of Oslo, Norway, 309 p.
- Balk H. & Lindem T., 2000. Improved fish detection in data from split-beam sonar. *Aquat. Living Res.* 13(5): 297–303.
- Balk H. & Lindem T., 2014. Sonar4 and Sonar5-Pro post processing systems, Operator manual version 6.0.3, 464 p.
- Belcher E., Hanot W. & Burch J., 2002. Dual-frequency identification sonar (DID-SON). In: *Underwater technology. Proceedings of the 2002 International Symposium on Underwater Technology*, pp. 187–192.
- Bevelhimer M.S. & Adams S.M., 1993. A bioenergetics analysis of diel vertical migration by kokanee salmon, *Oncorhynchus nerka*. *Can. J. Fish. Aquat. Sci.* 50(11): 2336–2349.
- Binder T.R. & McDonald D.G., 2007. The role of dermal photoreceptors during the sea lamprey (*Petromyzon marinus*) spawning migration. *Can. J. Fish. Aquat. Sci.* 194(11): 921–928.
- Bohl E., 1980. Diel pattern of pelagic distribution and feeding in planktivorous fish. *Oecologia* 44(3): 368–375.
- Bridger C.J. & Booth R.K., 2003. The effects of biotelemetry transmitter presence and attachment procedures on fish physiology and behavior. *Rev. Fish. Sci.* 11(3): 13–34.
- Brosse S., Dauba F., Oberdorff T. & Lek S., 1999a. Influence of some topographical variables on the spatial distribution of lake fish during summer stratification. *Arch. Hydrobiol.* 145(3): 359–371.
- Brosse S., Lek S. & Dauba F., 1999b. Predicting fish distribution in a mesotrophic lake by hydroacoustic survey and artificial neural networks. *Limnol. Oceanogr.* 44(5): 1293–1303.
- Busch S. & Mehner T., 2012. Size-dependent patterns of diel vertical migration: smaller fish may benefit from faster ascent. *Behav. Ecol.* 23(1): 210–217.
- Caroffino D.C., Sutton T.M. & Lindberg M. S., 2009. Abundance and movement patterns of age-0 juvenile lake sturgeon in the Peshtigo River, Wisconsin. *Environ. Biol. Fish.* 86(3): 411–422.
- Caswell N.M., Peterson D.L., Manny B.A. & Kennedy G.W., 2004. Spawning by lake sturgeon (*Acipenser fulvescens*) in the Detroit River. *J. Appl. Ichthyol.* 20(1): 1–6.
- Čech M. & Kubečka J., 2002. Sinusoidal cycling swimming pattern of reservoir fishes. *J. Fish Biol.* 61(2): 456–471.
- CEN, 2005. EN 14 757, CEN TC 230, Water quality – Sampling of fish with multi-mesh gillnets. European Committee for Standardization, Brussels.
- Cerri R.D., 1983. The effect of light intensity on predator and prey behaviour in cyprinid fish: factors that influence prey risk. *Anim. Behav.* 31(3): 736–742.
- Clark C.W. & Levy D.A., 1988. Diel vertical migrations by juvenile sockeye salmon and the antipredation window. *Am. Nat.* 131(2): 271–290.
- Cooke S.J., Hinch S.G., Lucas M.C. & Lutcavage M., 2012. Biotelemetry and biologging. In: *Fisheries techniques*,

- third edition.* (A.V. Zale, D.L. Parrish, T. M. Sutton, Eds.), pp. 819–881.
- Cooke S.J., Midwood J.D., Thiem J.D., Klimley P., Lucas M.C., Thorstad E.B., ... & Ebner B.C., 2013. Tracking animals in freshwater with electronic tags: past, present and future. *Anim. Biotelem.* 1(5): 1–19.
- Cooke S.J., Martins E.G., Struthers D.P., Gutowsky L.F., Power M., Doka S.E., ... & Krueger C.C., 2016. A moving target—incorporating knowledge of the spatial ecology of fish into the assessment and management of freshwater fish populations. *Environ. Monit. Assess.* 188(4): 1–18.
- Diehl S., 1988. Foraging efficiency of three freshwater fishes: effects of structural complexity and light. *Oikos* 53(2): 207–214.
- Djemali I., Toujani R. & Guillard J., 2009. Hydroacoustic fish biomass assessment in man-made lakes in Tunisia: horizontal beaming importance and diel effect. *Aquat. Ecol.* 43(4): 1121–1131.
- Dohet A. & Hoffmann L., 1995. Seasonal succession and spatial distribution of the zooplankton community in the reservoir of Esch-sur-Sûre (Luxembourg). *Belg. J. Zool.* 125: 109–123.
- Drašík V., Kubečka J., Čech M., Frouzová J., Říha M., Jůza T., ... & Vašek M., 2009. Hydroacoustic estimates of fish stocks in temperate reservoirs: day or night surveys? *Aquat. Living Resour.* 22(1): 69–77.
- Drašík V., Guillard J., Godlewská M., Claburn P., Hateley J., Kubečka J., Morrissey E., Winfield I.J., 2014. Inter-calibration of different hydroacoustic systems for the assessment of fish populations in deep lakes and reservoirs: towards a method for lake fish monitoring within the WFD. Oral Communication—ECOFIL. 8–11 September 2014, Ceske Budejovice, Czech Republic.
- Duncan A. & Kubečka J., 1993. *Hydroacoustic Methods of Fish Surveys.* National Rivers Authority, R&D Note 196, 136 p.
- Duncan A. & Kubečka J., 1995. Land/water ecotone effects in reservoirs on the fish fauna. *Hydrobiologia* 105: 11–30.
- Dunlop E.S., Milne S.W., Ridgway M.S., Condiotty J. & Higginbottom I., 2010. *In situ* swimming behavior of lake trout observed using integrated multibeam acoustics and biotelemetry. *Trans. Am. Fish. Soc.* 139(2): 420–432.
- Ebner B.C. & Thiem J.D., 2009. Monitoring by telemetry reveals differences in movement and survival following hatchery or wild rearing of an endangered fish. *Mar. Fresh. Res.* 60(1): 45–57.
- Eklov P. & VanKooten T 2001. Facilitation among piscivorous predators: effects of prey habitat use. *Ecology* 82(9), 2486–2494.
- Emmrich M., Winfield I.J., Guillard J., Rustadbakken A., Vergès C., Volta P., Jeppesen E., Lauridsen T., Holmgren K., Argillier C. & Mehner T., 2012. Strong correspondence between gillnet catch per unit effort and hydroacoustically derived fish biomass in stratified lakes. *Freshw. Biol.* 57(12): 2436–2448.
- Eriksson L.O., 1978. Nocturnalism versus diurnalism: dualism within fish individuals. In: *Rhythmic activity of fishes* (J.E. Thorpe, Ed.), pp. 69–89.
- Fernandes-Rosado M.J. & Lucena J., 2001. Space-time heterogeneities of the zooplankton distribution in La Concepcion reservoir (Istan, Malaga; Spain). *Hydrobiologia* 455(1–3): 157–170.
- Fernandes-Rosado M.J., Lucena K. & Niell F.X., 1994. Space-time heterogeneity of the chlorophyll-a distribution in La Concepcion reservoir (Istan, Malaga). Representative models. *Arch. Hydrobiol.* 129(3): 311–325.

- Fernando C.H. & Holčík J., 1991. Fish in reservoirs. *Intern. Revue Hydrobiol. Hydrogr.* 76(2): 149–167.
- Foote K., Knudsen H., Vestnes G., MacLennan D. & Simmonds E., 1987. Calibration of acoustic instruments for fish density estimation. *ICES Cooperative Report*. 144: 1–70.
- Frouzová J., Kubečka J., Balk H. & Frouz J., 2005. Target strength of some European fish species and its dependence on fish body parameters. *Fish. Res.* 75(1): 86–96.
- Fry F.E.J., 1971. The effect of environmental factors on the physiology of fish. *Fish physiol.* 6: 1–98.
- Gaudreau N. & Boisclair D., 1998. The influence of spatial heterogeneity on the study of fish horizontal daily migration. *Fish. Res.* 35(1): 65–73.
- Gillet C., 2001. Le déroulement de la fraie des principaux poissons lacustres. Dans: *Gestion piscicole des grands plans d'eau* (D. Gerdeaux Ed.), pp. 241–281.
- Gilliam J.F. & Fraser D.F., 2001. Movement in corridors: enhancement by predation threat, disturbance, and habitat structure. *Ecology* 82(1): 258–273.
- Gjelland K.O. & Hedger R.D., 2013. Environmental influence on transmitter detection probability in biotelemetry: developing a general model of acoustic transmission. *Methods Ecol. Evol.* 4(7): 665–674.
- Gliwicz Z.M., Slon J. & Szyrak I., 2006. Trading safety for food: evidence from gut contents in roach and bleak captured at different distances offshore from their daytime littoral refuge. *Freshw. Biol.* 51(5): 823–839.
- Godlewska M., Colon M., Jozwik A., Guillard J., 2011. How pulse lengths impact fish stock estimations during hydroacoustic measurements at 70 kHz. *Aquat. Living Res.* 24(1): 71–78.
- Godlewska M., Frouzová J., Kubečka J., Wisniewski W. & Szlakowski J., 2012. Comparison of hydroacoustic estimates with fish census in shallow Malta Reservoir—which TS/L regression to use in horizontal beam applications? *Fish. Res.* 123: 90–97.
- Grimardias D., Guillard J. & Cattaneo F., 2017. Drawdown flushing of a hydroelectric reservoir on the Rhône River: impacts on the fish community and implications for the sediment management of large dams. *J. Environ. Manage.*, in press.
- Guillard J. & Vergès C., 2007. The Repeatability of Fish Biomass and Size Distribution Estimates obtained by Hydroacoustic Surveys Using Various Survey Designs and Statistical Analyses. *Int. Rev. Hydrobiol.* 92(6): 605–617.
- Guillard J., Perga M.E., Colon M. & Angeli N., 2006. Hydroacoustic assessment of young-of-year perch, *Perca fluviatilis*, population dynamics in an oligotrophic lake (Lake Annecy, France). *Fish. Manage. Ecol.* 13(5): 319–327.
- Heupel M.R., Semmens J.M. & Hobday A. J., 2006. Automated acoustic tracking of aquatic animals: scales, design and deployment of listening station arrays. *Mar. Freshw. Res.* 57(1): 1–13.
- Horppila J., Ruuhijarvi J., Rask M., Karppinen C., Nyberg K. & Olin M., 2000. Seasonal changes in the diets and relative abundances of perch and roach in the littoral and pelagic zones of a large lake. *J. Fish Biol.* 56(1): 51–72.
- Hussey N.E., Kessel S.T., Aarestrup K., Cooke S.J., Cowley P.D., Fisk A.T., ... & Flemming J.E.M., 2015. Aquatic animal telemetry: a panoramic window into the underwater world. *Science* 348(6240): 1255642.
- Imbrock F., Appenzeller A. & Eckmann R., 1996. Diel and seasonal distribution of

- perch in Lake Constance: a hydro-acoustic study and *in situ* observations. *J. Fish Biol.* 49(1): 1–13.
- Jacobsen L., Berg S., Jepsen N. & Skov C., 2004. Does roach behaviour differ between shallow lakes of different environmental state? *J. Fish Biol.* 65 (1): 135–147.
- Järvalt A., Krause T. & Palm A., 2005. Diel migration and spatial distribution of fish in a small stratified lake. *Hydrobiologia* 547: 197–203.
- Jurajda P. & Regenda J., 2004. Littoral 0+ fish assemblages in three reservoirs of the Nove Mlýny dam (Czech Republic). *Czech J. Anim. Sci.* 49(10): 450–457.
- Kahl U. & Radke R.J., 2006. Habitat and food resource use of perch and roach in a deep mesotrophic reservoir: enough space to avoid competition? *Ecol. Freshw. Fish.* 15(1): 48–56.
- Kessel S.T., Cooke S.J., Heupel M.R., Hussey N.E., Simpfendorfer C.A., Vagle S. & Fisk A.T., 2014. A review of detection range testing in aquatic passive acoustic telemetry studies. *Rev. Fish. Biol. Fish.* 24(1): 199–218.
- Kessel S.T., Hussey N.E., Webber D.M., Gruber S.H., Young J.M., Smale M.J. & Fisk A.T., 2015. Close proximity detection interference with acoustic telemetry: the importance of considering tag power output in low ambient noise environments. *Anim. Biotelem.* 3(1): 5.
- Klimley A.P., Voegeli F., Beavers S.C. & Le Boeuf B.J., 1998. Automated listening stations for tagged marine fishes. *Marine Technology Society. Mar. Technol. Soc. J.* 32(1): 94.
- Knudsen F.R. & Sægrov H., 2002. Benefits from horizontal beaming during acoustic survey: application to three Norwegian lakes. *Fish. Res.* 56(2): 205–211.
- Kratochvil M., Čech M., Vašek M., Kubečka J., Hejzlar J., Matěna J., ... & Jarolím S. E.D.A., 2010. Diel vertical migrations of age 0+ percids in a shallow, well-mixed reservoir. *J. Limnol.* 69(2): 305–310.
- Kubečka J., 1993. Night inshore migration and capture of adult fish by shore seining. *Aquacult. Fish. Manage.* 24(5): 685–689.
- Kubečka J. & Duncan A., 1998. Diurnal changes of fish behaviour in a lowland river monitored by a dual-beam echosounder. *Fish. Res.* 35(1): 55–63.
- Kubečka J. & Wittingerova M., 1998. Horizontal beaming as a crucial component of acoustic fish stock assessment in freshwater reservoirs. *Fish. Res.* 35 (1): 99–106.
- Kubečka J., Frouzová J., Balk H., Čech M., Drašík V. & Prchalová M., 2009. Regressions for conversion between target strength and fish length in horizontal acoustic surveys. In: *Underwater acoustic measurements, Technologies & Results* (J.S. Papadakis, L. Bjorno, Eds.), pp. 1039–1044. Heraklion, Greece: Foundation for Research & Technology.
- Lewin W.C., Okun N. & Mehner T., 2004. Determinants of the distribution of juvenile fish in the littoral area of a shallow lake. *Freshw. Biol.* 49(4): 410–424.
- Love R.H., 1971. Dorsal-aspect target strength of an individual fish. *J. Acoust. Soc. Am.* 49(3B): 816–823.
- Love R.H., 1977. Target strength of an individual fish from any aspect. *J. Acoust. Soc. Am.* 62(6): 1397–1403.
- Lucas M.C. & Baras E., 2000. Methods for studying spatial behaviour of freshwater fishes in the natural environment. *Fish Fish.* 1(4): 283–316.
- Lucas M.C., Baras E., Thom T.J., Duncan A. & Slavič O., 2001. *Migration of freshwater fishes*. Oxford: Blackwell Science, 420 p.
- Lyons J. & Lucas M.C., 2002. The combined use of acoustic tracking and echosounding to investigate the movement and distribution of common bream (*Abramis brama*) in the River Trent, England. *Hydrobiologia* 483(1–3): 265–273.

- McCauley M.M., Cerrato R.M., Sclafani M. & Frisk M.G., 2014. Diel behavior in white perch revealed using acoustic telemetry. *Trans. Am. Fish. Soc.* 143(5): 1330–1340.
- McGrath K.J., Ault S., Reid K., Stanley D. & Voegeli F., 2003. Development of hydrosonic telemetry technologies suitable for tracking American eel movements in the vicinity of a large hydroelectric project. *Am. Fish. S. S.* 33: 329–341.
- Mehner T., 2012. Diel vertical migration of freshwater fishes-proximate triggers, ultimate causes and research perspectives. *Freshw. Biol.* 57(7): 1342–1359.
- Mehner T. & Kasprzak P., 2011. Partial diel vertical migrations in pelagic fish. *J. Anim. Ecol.* 80(4): 761–770.
- Mehner T., Diekmann M., Bramick U. & Lemcke R., 2005. Composition of fish communities in German lakes as related to lake morphology, trophic state, shore structure and human-use intensity. *Freshw. Biol.* 50(1): 70–85.
- Mehner T., Busch S., Helland I.P., Emmrich M. & Freyhof J., 2010. Temperature-related nocturnal vertical segregation of coexisting coregonids. *Ecol. Freshw. Fish.* 19(3): 408–419.
- Muška M., Tušer M., Frouzová J., Drašík V., Čech M., Jůza T., ... & Říha M., 2013. To migrate, or not to migrate: partial diel horizontal migration of fish in a temperate freshwater reservoir. *Hydrobiologia* 707(1): 17–28.
- Neverman D. & Wurtsbaugh W.A., 1994. The thermoregulatory function of diel vertical migration for a juvenile fish, *Cottus extensus*. *Oecologia* 98(3–4): 247–256.
- Ohlberger J., Staaks G., Petzoldt T., Mehner T. & Holker F., 2008. Physiological specialization by thermal adaptation drives ecological divergence in a sympatric fish species pair. *Evol. Ecol. Res.* 10(8): 1173–1185.
- Ottera H. & Skilbrei O.T., 2016. Influence of depth, time and human activity on detection rate of acoustic tags: a case study on two fish farms. *J. Fish Biol.* 88(3): 1229–1235.
- Parker A.D., Stepie, C.A., Sepulveda-Villet O.J., Ruehl C.B. & Uzarski D.G., 2009. The interplay of morphology, habitat, resource use, and genetic relationships in young yellow perch. *Trans. Am. Fish. Soc.* 138(4): 899–914.
- Parkinson E.A., Rieman B.E. & Rudstam L.G., 1994. Comparison of acoustic and trawl methods for estimating density and age composition of Kokane. *Trans. Am. Fish. Soc.* 123(6): 841–854.
- Persson L., 1986. Temperature-induced shift in foraging ability in two fish species, roach (*Rutilus rutilus*) and perch (*Perca fluviatilis*): implications for coexistence between poikilotherms. *J. Animal. Ecol.* 55(3): 829–839.
- Pont D. & Amrani J., 1990. The effects of selective fish predation on the horizontal distribution of pelagic Cladocera in a southern French reservoir. *Hydrobiologia* 207(1): 259–267.
- Portner H.O. & Farrell A.P., 2008. Physiology and climate change. *Science* 322(5902): 690–692.
- Prado I.G. & Pompeu P.S., 2014. Vertical and seasonal distribution of fish in Três Marias reservoir. *Lake Reser. Manage.* 30(4): 393–404.
- Prchalová M., Drašík V., Kubečka J., Sricharoendham B., Schiemer F. & Vijverberg J., 2003. Acoustic study of fish and invertebrate behavior in a tropical reservoir. *Aquat. Living Res.* 16(3): 325–331.
- Prchalová M., Kubečka J., Vašek M., Peterka J., Sed'a J., Jiiza T., .. & Čech M., 2008. Distribution patterns of fishes in a canyon-shaped reservoir. *J. Fish Biol.* 73(1): 54–78.

- R Development Core Team, 2016. R version 3.3.1. R Project for Statistical Computing. Vienna, Austria. [www.r-project.org](http://www.r-project.org).
- Radke R.J. & Gaupisch A., 2005. Effects of phytoplankton-induced turbidity on predation success of piscivorous Eurasian perch (*Perca fluviatilis*): possible implications for fish community structure in lakes. *Naturwissenschaften* 92(2): 91–94.
- Rakowitz G., Herold W., Fesl C., Keckeis H., Kubečka J. & Balk H., 2008. Two methods to improve the accuracy of target-strength estimates for horizontal beaming. *Fish. Res.* 9 (3): 324–331.
- Renard D. & Bez N., 2005. RGeoS: geostatistical package. R package. Version 2.1. Fontainebleau, France: Centre de Géostatistique, École des Mines de Paris.
- Říha M., Kubečka J., Prchalová M., Mrkvička T., Čech M., Drašík V., ... & Peterka J., 2011. The influence of diel period on fish assemblage in the unstructured littoral of reservoirs. *Fish. Manage. Ecol.* 18(4): 339–347.
- Říha M., Ricard D., Vašek M., Prchalová M., Mrkvička T., Jiiza T., ... & Peterka J., 2015. Patterns in diel habitat use of fish covering the littoral and pelagic zones in a reservoir. *Hydrobiologia* 747(1): 111–131.
- Romare P., Berg S., Lauridsen T. & Jeppesen E., 2003. Spatial and temporal distribution of fish and zooplankton in a shallow lake. *Freshw. Biol.* 48(8): 1353–1362.
- Roy R., 2014. Distribution spatiale et activité des poissons en milieu lacustre—Impacts des facteurs environnementaux à partir d’une approche multi-échelle. Application à la retenue des Bariousses. PhD thesis, Aix-Marseille. 224 p.
- Roy R., Beguin J., Argillier C., Tissot L., Smith F., Smedbol S. & De-Oliveira E., 2014. Testing the VEMCO Positioning System: spatial distribution of the probability of location and the positioning error in a reservoir. *Anim. Biotelem.* 2(1): 1–6.
- Rudstam L.G., Jech J.M., Parker-Stetter S. L., Horne J.K., Sullivan P.J. & Mason D. M., 2012. Fisheries acoustics. In: *Fisheries techniques, third edition* (A.V. Zale, D.L. Parrish, T.M. Sutton, Eds.), pp. 597–636.
- Savino J.F. & Stein R.A., 1989. Behavioural interactions between fish predators and their prey: effects of plant density. *Anim. behav.* 37: 311–321.
- Siler J.R., Foris W.J. & McInerney M.C., 1986. Spatial heterogeneity in fish parameters within a reservoir. In: *Reservoir Fisheries Management: Strategies for the 80’s* (G.E. Hall, M.J. Van Den Avyle, Eds), pp. 122–136.
- Simmonds E.J. & MacLennan D.N. 2005. *Fisheries Acoustics: Theory and Practice*. Oxford: Blackwell Science Ltd, 437 p.
- Smith F., 2013. Understanding HPE in the VPS Telemetry System. VEMCO Tutorials.
- Smith J.A., Baumgartner L.J., Suthers I.M. & Taylor M.D., 2011. Distribution and movement of a stocked freshwater fish: implications of a variable habitat volume for stocking programs. *Mar. Freshw. Res.* 62(11): 1342–1353.
- Straskraba M., 1998. Limnological differences between deep valley reservoirs and deep lakes. *Int. Rev. Hydrobiol.* 83: 1–12.
- Świerzowski A., Godlewska M. & Pottorak T., 2000. The relationship between the spatial distribution of fish, zooplankton and other environmental parameters in the Solina reservoir, Poland. *Aquat. Living Resour.* 13(5): 373–377.
- Tukey J.W., 1977. *Exploratory Data Analysis*. Reading, PA: Addison-Wesley, 688 p.

- Tušer M., Kubečka J., Frouzová J. & Jarolím O., 2009. Fish orientation along the longitudinal profile of the Rímov reservoir during daytime: Consequences for horizontal acoustic surveys. *Fish. Res.* 96(1): 23–29.
- Urabe J., 1989. Relative importance of temporal and spatial heterogeneity in the zooplankton community of an artificial reservoir. *Hydrobiologia* 184(1): 1–6.
- Urabe J., 1990. Stable horizontal variation in the zooplankton community structure of a reservoir maintained by predation and competition. *Limnol. Oceanogr.* 35 (8): 1703–1717.
- Vašek M., Kubečka J. & Sed'a J., 2003. Cyprinid predation on zooplankton along the longitudinal profile of a canyon-shaped reservoir. *Arch. Hydrobiol.* 156(4): 535–550.
- Vašek M., Kubečka J., Peterka J., Čech M., Draščík V., Hladík M., ... & Frouzová J., 2004. Longitudinal and Vertical Spatial Gradients in the Distribution of Fish within a Canyon-shaped Reservoir. *Int. Rev. Hydrobiol.* 89(4): 352–362.
- Vašek M., Kubečka J., Matěna J. & Sed'a J., 2006. Distribution and Diet of 0+ Fish within a Canyon-Shaped European Reservoir in Late Summer. *Int. Rev. Hydrobiol.* 91(2): 178–194.
- Vašek, M., Prchalová M., Říha M., Blabolil P., Čech M., Draščík V., ... & Peterka J., 2016. Fish community response to the longitudinal environmental gradient in Czech deep-valley reservoirs: implications for ecological monitoring and management. *Ecol. Indic.* 63: 219–230.
- Winfield I.J., Fletcher J.M., James J.B. & Bean C.W., 2009. Assessment of fish populations in still waters using hydroacoustics and survey gill netting: experiences with Arctic charr (*Salvelinus alpinus*) in the UK. *Fish. Res.* 96(1): 30–38.
- Wurtsbaugh W.A. & Neverman D., 1988. Post-feeding thermotaxis and daily vertical migration in a larval fish. *Nature* 333(6176): 846–848.
- Ye S., Lian Y., Godlewská M., Liu J. & Li Z., 2013. Day-night differences in hydroacoustic estimates of fish abundance and distribution in Lake Laojianghe, China. *J. Appl. Ichthyol.* 29(6): 1423–1429.
- Yule D.L., 2000. Comparison of horizontal acoustic and purse-seine estimates of salmonid densities and sizes in eleven Wyoming waters. *North Am. J. Fish. Manage.* 20(3): 759–775.
- Yule D., Evrard L.M., Cachera S., Colon M. & Guillard J., 2013. Comparing two fish sampling standards over time: largely congruent results but with caveats. *Freshw. Biol.* 58(10): 2074–2088.
- Zamora L. & Moreno-Amich R., 2002. Quantifying the activity and movement of perch in a temperate lake by integrating acoustic telemetry and a geographic information system. *Hydrobiologia* 483: 209–218.